

Effect of short-term exposures to traffic air pollution and temperature on heart rate among healthy female students during commmuting

by V D

Submission date: 12-Apr-2023 09:25AM (UTC+0700)

Submission ID: 2062134157

File name: t-term_exposures_to_traffic_air_pollution_and_temperature_on.pdf (3.01M)

Word count: 7217

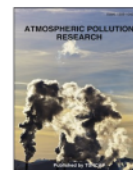
Character count: 34213



9

Contents lists available at ScienceDirect

Atmospheric Pollution Research

journal homepage: www.elsevier.com/locate/apr

10

Effect of short-term exposures to traffic air pollution and temperature on heart rate among healthy female students during commuting

Mandroy Pangaribuan^{a,b,*}, Diana Vanda Daturara Doda^{b,c}, Grace E.C. Korompis^b,
Woodford B.S. Joseph^b, Ribka Wowor^b, Veronika Simangunsong^d

^a Laboratory of Environmental and Occupational Health and Safety, Faculty of Public Health, Universitas Sam Ratulangi, Indonesia

^b Faculty of Public Health, Universitas Sam Ratulangi, Indonesia

^c Faculty of Medicine, Universitas Sam Ratulangi, Indonesia

^d Environmental Health and Disease Control Engineering Center, Ministry of Health, Indonesia



ARTICLE INFO

Keywords:

Exposures
Traffic air pollution
Temperature
Interaction effect
Heart rate

ABSTRACT

Findings showed that studies on the interaction effect of air pollution and temperature on biomarkers of cardiovascular health in a micro-environment (MiE) are still very limited. Therefore, this study aims to determine the interaction effect between air pollution, temperature, and biomarkers of cardiovascular outcomes, including Heart Rate/HR, Systolic Blood Pressure/SBP, and Diastolic Blood Pressure/DBP among healthy female students during commuting in a MiE, namely a public transport. A ParticleScan™ CR device (IQAir Group, Inc., Swiss) was used to measure Particulate Matter (PM). Temperature, Relative Humidity (RH), Nitrogen Dioxide (NO₂), Nitric Oxide (NO), Carbon Monoxide (CO), and Total Volatile Organic Compounds (TVOC) were measured using a YES Plus LGA multi-sensor device (Critical Environment Technologies Canada, Inc., Canada). The heart rates of the samples were assessed using an electronic blood pressure monitor Mediatech (Shanzen ZhengKang Technology Co., Ltd., China), while the vital lung capacity was obtained with a portable spirometer MicroLab (Micro Direct, Inc., USA). The data obtained in this study were analyzed using multiple linear regression. Furthermore, normality tests were carried out for all dependent variables, namely HR, SBP, and DBP using the Kolmogorov-Smirnov test and the result shows normal distribution. The beta coefficients (β) were used to express the association between the data. For the single effects, PM_{≥ 0.3}, PM_{≥ 0.5}, PM_{≥ 0.7}, PM_{≥ 1.0}, PM_{≥ 2.0}, PM_{0.3-0.5}, PM_{0.5-0.7}, PM_{0.7-1.0}, PM_{1.0-2.0}, and PM_{2.0-5.0} NO₂, and TVOC had a significant negative effect on HR; NO affected SBP and DBP negatively; and NO₂ had a positive influence on DBP. For the interaction effects, PM_{≥ 0.3}, PM_{≥ 0.5}, PM_{≥ 0.7}, PM_{0.3-0.5}, PM_{0.5-0.7}, and PM_{0.7-1.0} had a significant positive interaction effect with temperature on HR.

1. Introduction

For the past 30 years, scientists have been linking air pollution with health outcomes (Cao et al., 2018; Ezzati and Dockery, 2009; Miller et al., 2007; Roh et al., 2020; Stafoggia et al., 2022). However, most of the studies carried out focused on a macro-environment (MaE) scale, such as country (Cao et al., 2018; Roh et al., 2020) and long-term effects (Miller et al., 2007; Stafoggia et al., 2022), which led to difficulty in taking necessary corrective policy required at such scale (Reis et al., 2022). Previous reports also showed that studies on a micro-environment (MiE) scale, such as public transportation is still very limited (Peters et al., 2004; Setton et al., 2011) and the results are

still unclear due to differences in the methodology, samples, and dependent variable (Biel et al., 2020; Zhang et al., 2022). People can spend a lot of time in a particular MiE, including kitchen, office, and vehicle. A previous study revealed that exposure to specific MiE-related air pollution is two times greater than that of MaE (Peters et al., 2004). Furthermore, studies on exposure to MiE are urgently needed to implement the immediate necessary policy (Caplin et al., 2019; Heal et al., 2012). Biases that arise from a MiE-related study are fewer than from MaE. For example, the use of real-time personal exposure in a MiE-related study (Biel et al., 2020) versus the application of ambient exposure (Miller et al., 2007) and modeling (Roh et al., 2020) in others associated with MaE.

Peer review under responsibility of Turkish National Committee for Air Pollution Research and Control.

* Corresponding author. Laboratory of Environmental and Occupational Health and Safety, Faculty of Public Health, Universitas Sam Ratulangi, Indonesia.

E-mail address: mandroypangaribuan@unsrat.ac.id (M. Pangaribuan).

<https://doi.org/10.1016/j.apr.2022.101631>

Received 3 September 2022; Received in revised form 7 December 2022; Accepted 8 December 2022

Available online 9 December 2022

1309-1042/© 2022 Turkish National Committee for Air Pollution Research and Control. Production and hosting by Elsevier B.V. All rights reserved.

One of the important MiE is public transportation, where people can spend hours traveling. In Indonesia, the most common medium of public transportation is *angkot*, a minibus with 9–14 seats. Most of the *angkots* in the country are obsolete since there is no strict regulation regarding their operating age limit. Previous studies revealed that the older the vehicle, the more pollution it emits (Tartakovsky et al., 2013; Wu et al., 2022). Furthermore, most *angkots* are also not equipped with air conditioning, hence, all the windows and doors are always open during commuting. Previous studies revealed that air pollution received by passengers is higher when vehicle windows are opened (Lim et al., 2021; Tartakovsky et al., 2013). Several studies showed that most of the pollution in MaE comes from the transportation MiE (Suarez-Bertoa et al., 2021; Walsh, 2008) and their number is expected to continue increasing rapidly in the future (Walsh, 2008). This is also exacerbated by long traffic jams that allow passengers to be exposed to more pollution for a longer period (Lipfert and Wyzga, 2008).

Apart from air pollution, passengers in *angkot* are also exposed to temperature (Grundstein et al., 2009; McLaren et al., 2005). Furthermore, literature studies show that the adverse effect of air pollutants on human health is dominated by cardiovascular outcomes (Biel et al., 2020; Miller et al., 2007; Peters et al., 2004; Stafoggia et al., 2022). Previous studies have also reported that temperature can affect cardiovascular health. Kupcikova et al. (2021) stated that temperature was associated with alterations in cardiac autonomic function in healthy adults affected by traffic-related air pollution (Kupcikova et al., 2021). Although air pollution and temperature have independent effect on cardiovascular health, studies investigating the interaction effect between both of them in a MiE is still very limited. Therefore, this study aims to investigate the interaction effect between air pollution, temperature, and biomarkers of cardiovascular outcomes, including Heart Rate/HR, Systolic Blood Pressure/SBP, and Diastolic Blood Pressure/DBP (Pangaribuan et al., 2019) in healthy female students. Females were selected as the samples because they are more susceptible to adverse effects of pollutants than males (Hong et al., 2005; Phung et al., 2016).

2. Material and methods

2.1. Study design and samples

An electronic invitation flyer was distributed at the Faculty of Public Health, Sam Ratulangi University, Indonesia to recruit samples. The inclusion criteria in this study were healthy female students with no history of cardiovascular and asthma disorder, not a smoker, and having received three doses of the Covid-19 vaccine. A total of 4 students who met the criteria were then selected to participate in the process. Before the start of this study, a 1-day training on all the instruments was carried out, as well as the distribution of informed consent. One *angkot* with a typical age of 30 years was used to transport all the samples and instruments. There were a total of 6 passengers in the *angkot*, namely 4 students, 1 researcher, and 1 driver. This study was carried out on August 4, 2022, from 08:00 to 16:00 (GMT+8). During the study, the passenger door and windows of the *angkot* were conditioned as usual (in the open state). A sample interval of 2 min was used for air pollution, temperature, RH, HR, SBP, and DBP.

Particulate Matter (PM) was measured using a ParticleScan™ CR device (IQAir Group, Inc., Swiss). ParticleScan™ CR uses a laser diode light source and collection optics for particle detection. Particles scatter the light from the laser diode beam in the direction of the collection optics. The collection optics focus the light onto a photodiode that

converts the bursts of light into electrical impulses. The pulse height is proportional to the particle size. Impulses are counted and their intensity is measured for particle sizing. The results are then displayed digitally for the specific size channel(s) and set measurement unit. ParticleScan™ CR has the following technical specification: minimum sensitivity of 0.3 μm, flow rate of 0.0028 cubic meters per minute, and sample time of 1–600 s. ParticleScan™ CR can measure cumulative size of PM (PM_{≥0.3}, PM_{≥0.5}, PM_{≥0.7}, PM_{≥1.0}, PM_{≥2.0}, and PM_{≥5.0}) and differential size of PM (PM_{0.3-0.5}, PM_{0.5-0.7}, PM_{0.7-1.0}, PM_{1.0-2.0}, and PM_{2.0-5.0}) at the same time. PM_{≥0.3} is defined as particulate matter with an aerodynamic diameter greater than and equal to 0.3 μm, and so on. PM_{0.3-0.5} is defined as particulate matter with an aerodynamic diameter between 0.3 and 0.5 μm, and so on.

Temperature, Relative Humidity (RH), Nitrogen Dioxide (NO₂), Nitric Oxide (NO), Carbon Monoxide (CO), and Total Volatile Organic Compounds (TVOC) were measured using a YES Plus LGA multi-sensor (Critical Environment Technologies Canada, Inc., Canada) device. The YES Plus LGA is a battery-powered, portable air quality monitor - information recording (data logging) instrument designed for intermittent or continuous operation. The standard range of its measurements is: temperature (−5.0 – 50 °C, 1 °C resolution), RH (0–100%, 1% resolution), NO₂ (0–5.0 ppm, 0.1 ppm resolution), NO (0–100 ppm, 0.1 ppm resolution), CO (0–50 ppm, 1 ppm resolution), TVOC (0–300 ppm, 1 ppb resolution), where ppm stands for parts per million and ppb stands for parts per billion.

2.2. Measurement of heart rate and vital lung capacity

The samples' heart rates were measured using an electronic blood pressure monitor Mediatech (Shanzen ZhengKang Technology Co., Ltd., China). It has the following technical specifications: measurement mode (oscillometric method), measurement range (40–195 times/min), and accuracy (±5%). The samples' vital lung capacities were measured using a portable spirometer MicroLab (Micro Direct, Inc., USA) which has measurement accuracy of ±3%. Four vital capacity indices of the lungs were measured in the process including Forced Expired Volume in 1 s (FEV₁), Forced Vital Capacity (FVC), Peak Expiratory Flow Rate (PEF), and Forced expiratory flow at 50% of exhaled volume (FEF₅₀).

2.3. Quality assurance

Calibration and “zero count” tests for all the instruments were carried out before the start of the process. The protocols for calibration and “zero count” tests for all the instruments are explained in the device manual books and elsewhere.

2.4. Statistical analysis

The data obtained in this study were analyzed using multiple linear regression. Normality tests were carried out for all dependent variables, including HR, SBP, and DBP using kolmogorov-smirnov test, and the result shows that the data are normally distributed. Subsequently, the beta coefficients (β) were used to express the association. All the statistical analyses were performed with IBM® SPSS® version 26 software. The following equation was then used to quantify the single effect of air pollution on HR, SBP, and DBP:

$$HR / SBP / DBP = \beta_0 + \beta_1 \text{ Air Pollution} + \varepsilon$$

The equation below was used to quantify the combined effect of air pollution and temperature on HR, SBP, and DBP:

$$HR / SBP / DBP = \beta_0 + \beta_1 \text{ Air Pollution} + \beta_2 \text{ Temperature} + \beta_3 \text{ Air Pollution} * \text{Temperature} + \varepsilon$$

Table 1
Characteristics of the samples.

Variables	Mean \pm SD
Age (years)	21 \pm 0.82
Body Mass Index/BMI (kg/m ²)	20.65 \pm 2.34
Heart Rate/HR (beats/minute)	84.34 \pm 11.36
Systolic Blood Pressure/SBP (mmHg)	117.56 \pm 20.68
Diastolic Blood Pressure/DBP (mmHg)	76.73 \pm 22.53

Where β_0 denotes the regression model intercept; β_1 , β_2 , and β_3 denote beta coefficients for each covariate; and ϵ denotes a residual error. Increase in the interquartile range (IQR) of each covariate was used to express increment in HR, SBP, and DBP.

3. Results and discussion

3.1. Characteristics and vital lung capacity indices of samples

Table 1 shows the characteristics of the samples, namely four female students with a mean age and Body Mass Index (BMI) of 21 \pm 0.82 years and 20.65 \pm 2.34, respectively. According to the World Health Organization (WHO) classification, they had normal weight (WHO, 2010). The samples used in this study were healthy with no history of cardiovascular and asthma disorders as well as smoking to minimize biases. They had also received three doses of the Covid-19 vaccine and passed the coronavirus symptoms screening before the start of the process. The mean HR, SBP, and DBP of the samples were 84.34 \pm 11.36 beats/minute, 117.56 \pm 20.68 mmHg, and 76.73 \pm 22.53 mmHg, respectively, and the box plot is presented in Fig. 13.

This study was carried out in one day on August 4, 2022, starting at 08:00 to 16:00 (GMT+8). A total of 8 h were used for measurement with a 1-h lunch break. Fig. 1 shows the study location and GPS tracking of the *angkot* travel routes from the beginning to the end of the experiment. The study began with the measurement of the samples' vital lung capacity. The process was then repeated shortly after the end of the study

at 16.00 (GMT+8). For each period, 2 measurements of vital lung capacity were taken and the best results were selected for further analysis.

Table 2 shows the results of measuring the vital lung capacity indices of the samples. Compared to the initial measurement, the indices obtained in the second round decreased by 6.32% (2.295 L–2.15 L), 8.03% (2.4275 L–2.2325 L), 11.72% (215.5 L/minute to 190.25 L/minute), and 2.64% (3.0325 L/second to 2.9525 L/second) for FEV₁, FVC, PEF, and FEF₅₀, respectively. This is in line with a previous study that found an increase in the interquartile range of PM₁₀ associated with a 5.1% decrease in FEV₁ from 95% CI 2.5%–7.7% as well as a 3.7% decrement in FVC from 95% CI 1.8%–5.5% among COPD women (Schikowski et al., 2005). Another study reported a 10 $\mu\text{g}/\text{m}^3$ increase in acute PM₁₀ exposure, which was related to a -0.19 l/min change in PEF (Edginton et al., 2021).

3.2. Characteristics of air pollution and meteorological variables

Figs. 2–12 show the plot of 2 min mean of air pollution and meteorological variables, while Table 3 presents the statistical description. The means of temperature and RH were 32.53 \pm 2.14 $^{\circ}\text{C}$ and 60.24 \pm 8.14%, respectively. The concentration means of PM ≥ 0.3 , PM ≥ 0.5 , PM

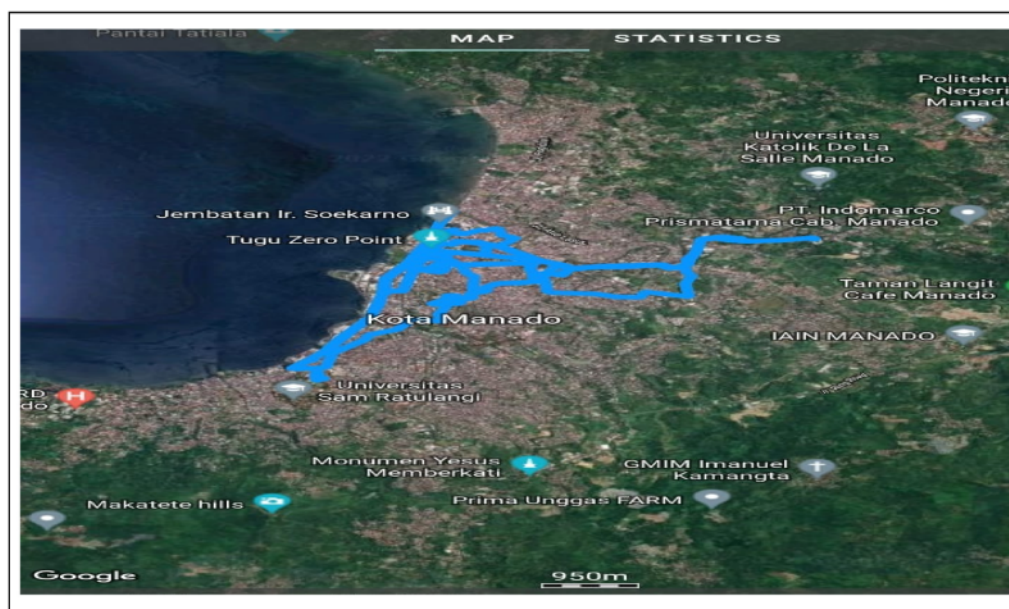
Table 2
Vital lung capacity indices of the samples.

Indices	Units	Min	Pred	Max	V ₁	V ₂
FEV ₁	Liter	2.0775	2.495	2.9425	2.295	2.15
FVC	Liter	2.265	2.7625	3.2625	2.4275	2.2325
PEF	Liter/minute	266	412	559	215.5	190.25
FEF ₅₀	Liter/second	NA	NA	NA	3.0325	2.9525

FEV₁: Forced Expired Volume in 1 s; FVC: Forced Vital Capacity; PEF: Peak Expiratory Flow Rate; FEF₅₀: Forced expiratory flow at 50% of exhaled volume. V₁: First measurement of vital lung capa

Min: the minimum value of the vital lung capacity index of Asian women; Pred: the predictive value/mean value of vital lung capacity index of Asian women; Max: the maximum value of the lung vital capacity index of Asian women.

NA: Not Available.



— GPS tracking of the *angkot* travel routes

Fig. 1. The study location and the GPS tracking of the *angkot* travel routes.

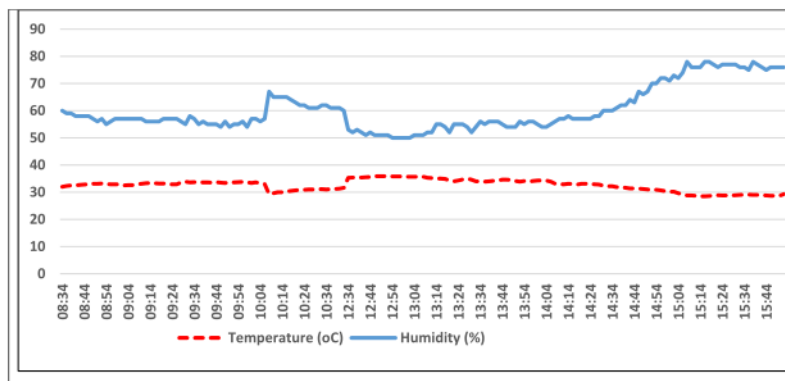


Fig. 2. Plot of 2-min mean temperature and RH.

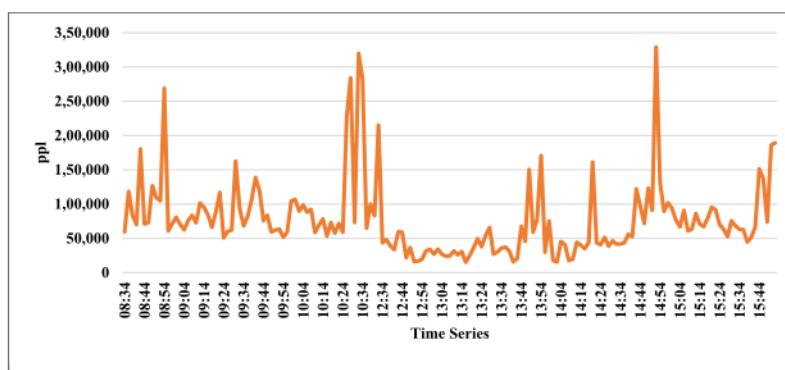


Fig. 3. Plot of 2-min mean $PM_{>0.3}$ concentration.

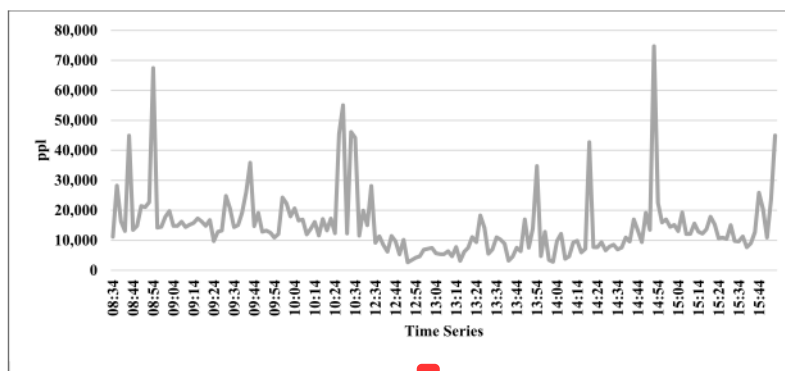


Fig. 4. Plot of 2-min mean $PM_{>0.5}$ concentration.

$PM_{>0.7}$, $PM_{>1.0}$, $PM_{>2.0}$, and $PM_{>5.0}$ were $77,476.67 \pm 55,496.61$, $14,943.66 \pm 10,972.04$, 6252.03 ± 4350.29 , 2158.98 ± 1689.12 , 656.71 ± 436.94 , and 12.97 ± 10.87 , respectively. Furthermore, the average level of $PM_{0.3-0.5}$, $PM_{0.5-0.7}$, $PM_{0.7-1.0}$, $PM_{1.0}$, and $PM_{2.0-5.0}$, were $62,533.01 \pm 45,287.67$, 8691.63 ± 6688.05 , 4093.05 ± 2804.07 , 1502.27 ± 1283.97 , and 643.7 ± 430.62 , respectively. All the PM air pollutants were measured in particles per liter/ppl. The mean concentration of NO_2 , CO, NO, and TVOC were 0.07 ± 0.04 , 11.31 ± 5.36 , 0.48 ± 0.23 , and 42.00 ± 6.85 , respectively. All the gas air pollutants were measured in parts per million/ppm, excluding TVOC in parts per billion/

ppb.

3.3. The single effect of air pollution on HR, SBP, and DBP

Table 4 shows the single effect of air pollution on HR, SBP, and DBP. For the PM pollutant $PM_{>0.3}$, $PM_{>0.5}$, $PM_{>0.7}$, $PM_{>1.0}$, $PM_{>2.0}$, $PM_{0.3-0.5}$, $PM_{0.5-0.7}$, $PM_{0.7-1.0}$, $PM_{1.0-2.0}$, and $PM_{2.0-5.0}$ had a significant effect on HR, where an increase of 1 IQR in the parameters led to a decrease in HR by 1.08, 0.96, 0.94, 1.06, 1.05, 1.04, 0.94, 0.80, 1.04, 1.05 beats/minute, respectively. Furthermore, for the gaseous

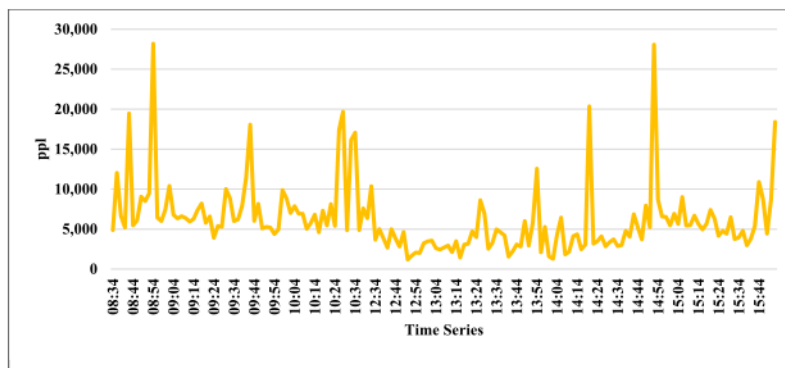


Fig. 5. Plot of 2-min mean $PM_{\geq 0.7}$ concentration.

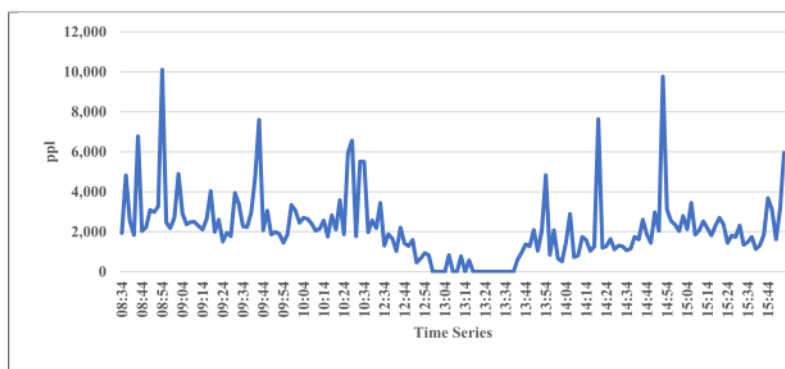


Fig. 6. Plot of 2-min mean $PM_{\geq 1.0}$ concentration.

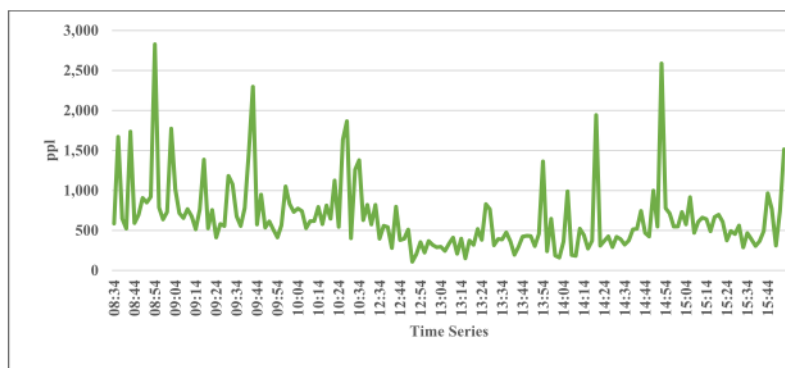


Fig. 7. Plot of 2-min mean $PM_{\geq 2.0}$ concentration.

pollutants, NO_2 and TVOC had a significant effect. An increase in NO_2 and TVOC concentration by 1 IQR led to a decrease in HR by 1.30 and 0.21 beats/minute, respectively.

This study results are inconsistent with previous studies that reported a positive effect of air pollution on HR (Dockery et al., 1999; Liu et al., 2014; Longo et al., 2008; Peters et al., 1999; Pope et al., 1999). However, this study is in line with several studies, where a negative influence was recorded (Angela et al., 2004; Gold et al., 2000). The small sample size used is likely not to detect a positive effect of air pollution on HR. It is also plausible that in certain vulnerable populations, particulate

pollution can lead to dysregulation of autonomic function, thereby reducing heart rate (Gold et al., 2000).

The results showed that SBP was only affected by the NO , where an increase of 1 IQR in NO concentration led to a decrease in SBP by 1.46 mmHg. This finding is inconsistent with previous studies, where a positive correlation was reported between them (Bruce et al., 2005; Kateryna et al., 2014). However, the results of this study are in line with Mette et al. (2012), where a doubling of NO_x exposure during 1- and 5-year periods preceding enrollment was associated with a decrease of 0.53 mmHg in SBP (Mette et al., 2012).

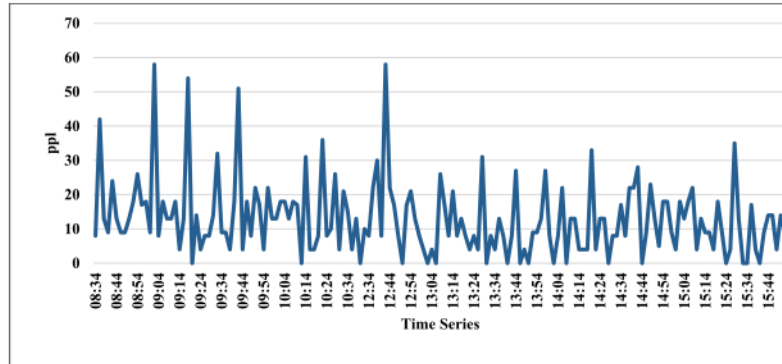
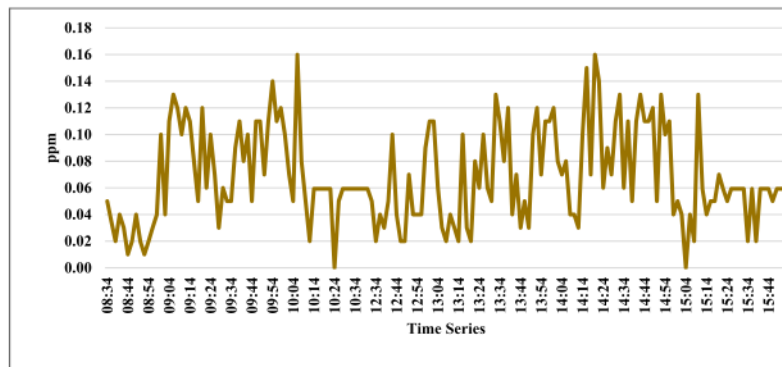
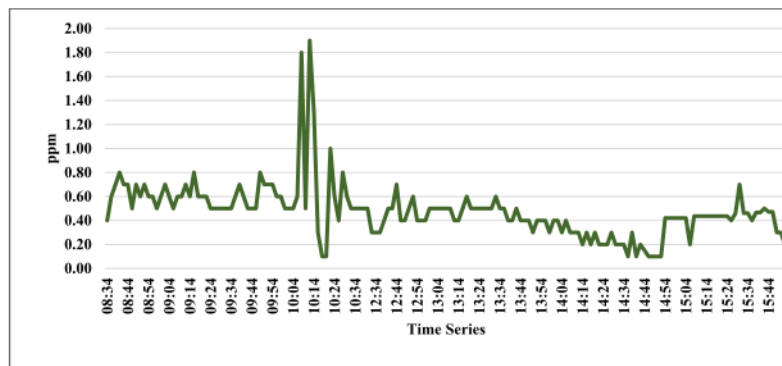
Fig. 8. Plot of 2-min mean $PM_{\geq 5.0}$ concentration.Fig. 9. Plot of 2-min mean NO_2 concentration.

Fig. 10. Plot of 2-min mean NO concentration.

The observation revealed that NO had a negative effect on DBP, where a 1 IQR decrease in NO concentration led to a reduction of 1.43 mmHg in DBP. This finding is inconsistent with Chen et al. (2015) who reported that an increase in NO_x by $20 \mu\text{g}/\text{m}^3$ was associated with an increase in diastolic blood pressure by 0.34 mmHg (Chen et al., 2015). The negative effect of this pollutant on SBP and DBP is likely due to a shift in the sympathovagal balance due to an increase in vagal tone. Another explanation is related to the effect of NO as a potent vasodilator that diffuses freely across membranes (Ibald-Mullis et al., 2004; Mette et al., 2012). The results showed that NO_2 had a positive effect on DBP,

where an increase in NO_2 concentration by 1 IQR led to an increase in DBP by 2.90 mmHg. This finding is in line with Li et al. (2020), where an increase in the level of the pollutant by $10 \mu\text{g}/\text{m}^3$ caused an increment in diastolic blood pressure by 0.797 mmHg (Li et al., 2020).

3.4. The interaction effect of air pollution and temperature on HR, SBP, and DBP

Table 5 shows the interaction effect of air pollution and temperature on HR, SBP, and DBP. For the PM pollutants, $PM_{\geq 0.3}$, $PM_{\geq 0.5}$, $PM_{\geq 0.7}$,

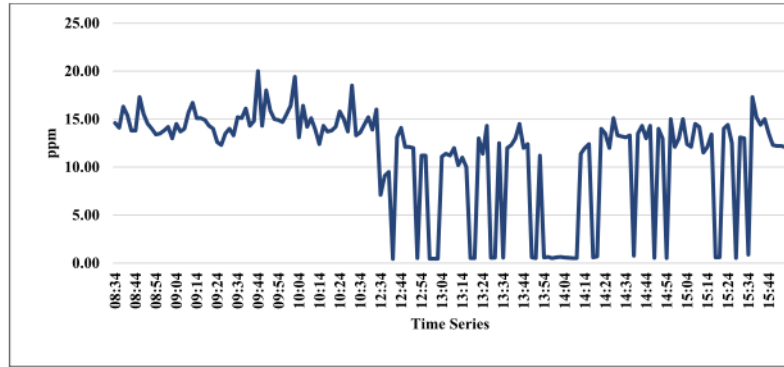


Fig. 11. Plot of 2-min mean CO concentration.

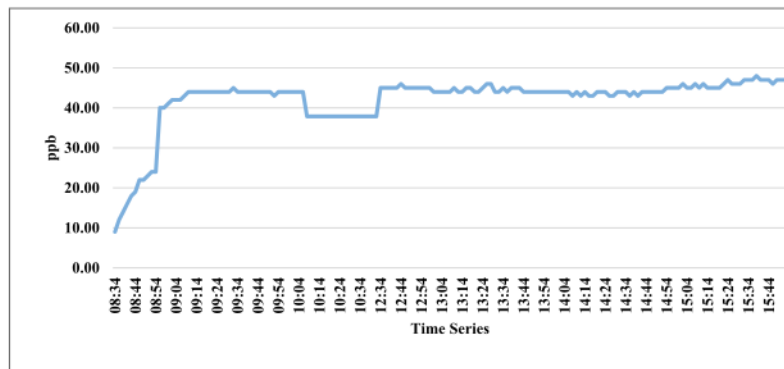


Fig. 12. Plot of 2-min mean TVOC concentration.

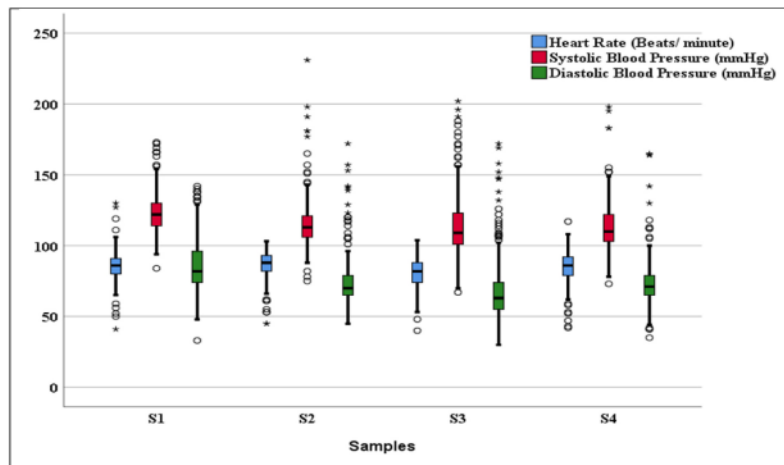


Fig. 13. The distribution of HR, SBP, DBP of the samples.

PM_{0.3-0.5}, PM_{0.5-0.7}, and PM_{0.7-1.0} had a significant interaction effect with temperature on heart rate. An increase in PM ≥ 0.3 , PM ≥ 0.5 , PM ≥ 0.7 , PM_{0.3-0.5}, PM_{0.5-0.7}, PM_{0.7-1.0}, and temperature by 1 IQR led to an increase in HR by 6.52, 4.52, 4.34, 6.75, 4.33, and 5.33 beats/minute, respectively. Previous studies have explored the single effect of temperature on heart rate. Madaniyazi et al. (2016) revealed that a 1 °C

increase in ambient temperature in a hot environment was associated with 0.133 beats/minute increment in HR compared to 0.063 beat/minute, which was obtained in a cold environment (Madaniyazi et al., 2016). A previous study revealed that heart rate increased significantly ($p < 0.05$) from 66.5 bpm pre-exposure to 106.0 bpm during exposure to dry heat. It was then concluded that being exposed to high ambient

Table 3
The statistical description of air pollutants and meteorological variables.

Tables	Mean ± SD	Minimum	Maximum	IQR
PM _{≥0.3} μm (ppl)	77,476.67 ± 55,496.61	15,164	329,071	47,118
PM _{≥0.5} μm (ppl)	14,943.66 ± 10,972.04	2635	74,722	8066,00
PM _{≥0.7} μm (ppl)	6252.03 ± 4350.29	1187	28,189	3205
PM _{≥1.0} μm (ppl)	2158.98 ± 1689.12	1.06	10,106	1335
PM _{≥2.0} μm (ppl)	656.71 ± 436.94	108	2828	375
PM _{≥5.0} μm (ppl)	12.97 ± 10.87	0	58	14
PM _{0.3-0.5} μm (ppl)	62,533.01 ± 45,287.67	12,028	273,565	38,553
PM _{0.5-0.7} μm (ppl)	8691.63 ± 6688.05	1448	46,644	4867
PM _{0.7-1.0} μm (ppl)	4093.05 ± 2804.07	724	18,314	1918
PM _{1.0-2.0} μm (ppl)	1502.27 ± 1283.97	826	7278	998
PM _{2.0-5.0} μm (ppl)	643.7 ± 430.62	108	2802	365
NO ₂ (ppm)	0.07 ± 0.04	0	0.16	0.06
CO (ppm)	11.31 ± 5.36	0.43	20	3
NO (ppm)	0.48 ± 0.23	0.10	1.90	0.20
TVOC (ppb)	42.00 ± 6.85	9	48.00	2
Temperature (°C)	32.53 ± 2.14	28.50	35.90	3
RH (%)	60.24 ± 8.14	50	78	8

temperature produces a significant increase in HR through an increase in sympathetic and a decrease in parasympathetic drive (Bruce-Low et al., 2006).

A previous study examined a modifying effect of PM_{2.5} on the relationship between temperature and HRV. The results showed that compared to the conditions of low PM_{2.5} (<12.9 μg/m³) levels, the effect of temperature on HRV was greater when air pollution was high, namely >12.9 μg/m³ (Ren et al., 2011). The findings in this study contribute to the literature that air pollution and temperature have a positive interaction effect on increasing heart rate. However, this study has several limitations, including the small number of samples, and the relatively short experimental period. This indicates that further studies with a large population and extended periods need to be carried out.

4. Conclusion

This study aims to determine the interaction effect of air pollution and temperature on cardiovascular health biomarker such as HR, SBP, and DBP. For the single effects, it was observed that PM_{≥0.3}, PM_{≥0.5}, PM_{≥0.7}, PM_{≥1.0}, PM_{≥2.0}, PM_{0.3-0.5}, PM_{0.5-0.7}, PM_{0.7-1.0}, PM_{1.0-2.0}, PM_{2.0-5.0}, NO₂, and TVOC had a significant negative influence on HR; NO affected SBP and DBP negative and NO₂ had a positive influence on DBP. For the interaction effects, PM_{≥0.3}, PM_{≥0.5}, PM_{≥0.7}, PM_{0.3-0.5}, PM_{0.5-0.7}, and PM_{0.7-1.0} had a significant positive interaction effect with

Table 4
The single effect of air pollutants and meteorological variables on HR, SBP, and DBP.

Variable	HR		SBP		DBP							
	β ₀	P-Value	β ₁	P-Value	β ₀	P-Value	β ₁	P-Value	β ₀	P-Value	β ₁	P-Value
PM _{≥0.3} μm (ppl)	86.12	<0.001***	-1.08	0.004**	116.12	<0.001***	0.89	0.198	75.37	<0.001***	0.85	0.268
PM _{≥0.5} μm (ppl)	86.11	<0.001***	-0.96	0.003**	116.36	<0.001***	0.64	0.276	75.70	<0.001***	0.56	0.389
PM _{≥0.7} μm (ppl)	86.18	<0.001***	-0.94	0.004**	116.28	<0.001***	0.65	0.271	75.57	<0.001***	0.60	0.357
PM _{≥1.0} μm (ppl)	86.05	<0.001***	-1.06	0.003**	116.31	<0.001***	0.77	0.229	75.52	<0.001***	0.75	0.282
PM _{≥2.0} μm (ppl)	86.18	<0.001***	-1.05	0.006**	116.48	<0.001***	0.61	0.376	75.70	<0.001***	0.59	0.432
PM _{≥5.0} μm (ppl)	83.89	<0.001***	0.49	0.392	117.46	<0.001***	0.11	0.919	77.05	<0.001***	-0.34	0.763
PM _{0.3-0.5} μm (ppl)	86.06	<0.001***	-1.04	0.005**	116.10	<0.001***	0.89	0.190	75.34	<0.001***	0.85	0.252
PM _{0.5-0.7} μm (ppl)	86.03	<0.001***	-0.94	0.003**	116.44	<0.001***	0.63	0.284	75.80	<0.001***	0.52	0.416
PM _{0.7-1.0} μm (ppl)	86.06	<0.001***	-0.80	0.008**	116.41	<0.001***	0.54	0.328	75.73	<0.001***	0.47	0.437
PM _{1.0-2.0} μm (ppl)	85.91	<0.001***	-1.04	0.002**	116.35	<0.001***	0.801	0.201	75.56	<0.001***	0.78	0.252
PM _{2.0-5.0} μm (ppl)	86.20	<0.001***	-1.05	0.005**	116.48	<0.001***	0.601	0.371	75.68	<0.001***	0.60	0.421
NO ₂ (ppm)	85.83	<0.001***	-1.30	0.08*	115.32	<0.001***	1.97	0.146	73.43	<0.001***	2.90	0.05**
CO (ppm)	84.82	<0.001***	0.13	0.609	117.06	<0.001***	0.13	0.769	77.20	<0.001***	-0.12	0.80
NO (ppm)	83.60	<0.001***	0.31	0.415	121.04	<0.001***	-1.46	0.035**	80.14	<0.001***	-1.43	0.06*
TVOC (ppb)	88.84	<0.001***	-0.21	0.09*	124.20	<0.001***	-0.32	0.180	80.62	<0.001***	-0.19	0.47
Temperature (°C)	52.35	<0.001***	2.95	<0.001***	143.21	<0.001***	-2.37	0.037**	100.01	<0.001***	-2.15	0.08*
RH (%)	97.02	<0.001***	-1.68	<0.001***	105.42	<0.001***	1.16	0.042**	65.86	<0.001***	1.44	0.09*

*: p-value <0.10 **: p-value <0.05 ***: p-value <0.01.

Table 5
The interaction effect of air pollutants and temperature on HR.

Variable	HR		β ₁	P-Value	β ₂	P-Value	β ₃	P-Value
	β ₀	P-Value						
PM _{≥0.3} *T	26.68	0.034**	21.01	0.005**	5.41	<0.001***	-19.90	0.004**
PM _{≥0.5} *T	32.66	0.009**	15.04	0.025**	4.88	<0.001***	-15.40	0.020**
PM _{≥0.7} *T	33.30	0.011**	13.39	0.046**	4.81	<0.001***	-13.86	0.037**
PM _{≥1.0} *T	44.82	<0.001***	8.02	0.276	3.71	0.001***	-8.53	0.241
PM _{≥2.0} *T	43.90	0.002**	7.49	0.373	3.85	0.003**	-7.78	0.325
PM _{≥5.0} *T	61.15	<0.001***	-10.21	0.284	2.12	0.027**	10.14	0.268
PM _{0.3-0.5} *T	26.46	0.033**	21.20	0.004**	5.42	<0.001***	-19.87	0.003**
PM _{0.5-0.7} *T	32.81	0.006**	15.87	0.017**	4.86	<0.001***	-16.40	0.014**
PM _{0.7-1.0} *T	31.56	0.015**	12.58	0.034**	4.98	<0.001***	-12.23	0.027**
PM _{1.0-2.0} *T	46.06	<0.001***	7.49	0.288	3.58	0.001***	-7.83	0.254
PM _{2.0-5.0} *T	43.34	0.002**	7.90	0.339	3.90	0.002**	-8.22	0.293
NO ₂ *T	42.39	0.006**	9.61	0.514	4.03	0.004**	-11.00	0.445
NO*T	34.20	0.065*	7.57	0.296	4.63	0.010**	-7.28	0.317
CO*T	40.63	0.012**	3.01	0.457	3.96	0.006**	-3.06	0.478
TVOC*T	123.32	0.266	-3.20	0.519	-3.20	0.755	8.07	0.548

*: p-value <0.10 **: p-value <0.05 ***: p-value <0.01.

temperature on HR.

Credit author statement

Mandroy Pangaribuan: conception and design of study, acquisition of data, Formal analysis and interpretation of data, drafting the manuscript, and revising the manuscript critically for important intellectual content. **Diana Vanda Daturara Doda:** conception and design of study, Formal analysis and interpretation of data, and revising the manuscript critically for important intellectual content. **Grace E.C. Korompis:** conception and design of study and revising the manuscript critically for important intellectual content. **Woodford B.S. Joseph:** acquisition of data and revising the manuscript critically for important intellectual content. **Ribka Wowor:** acquisition of data and revising the manuscript critically for important intellectual content. **Veronika Simangunsong:** revising the manuscript critically for important intellectual content.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Acknowledgments

This study was supported and funded by the Universitas Sam Ratulangi under research contract No: 157/UN12.13/LT/2022 and letter of assignment No: 662/UN12.13/LT/2022. The authors are grateful to Meivina Shalommita Putri Sumarauw, Annisa Puji Rahayu, and Yuniar Masloman for the technical support.

References

- Angela, I.-M. L. T. K., Annette, P., Joachim, H., Gabriele, W., Timo, L., Gintautas, B. G., K. W. Jeroen, de H., Gerard, H.M., ten, B.H., Juha, P., 2004. Effects of particulate air pollution on blood pressure and heart rate in subjects with cardiovascular disease. A multicenter approach. *Environmental Health Perspectives* 112 (3), 369–377. <https://doi.org/10.1289/ehp.6523>.
- Biel, R., Danieli, C., Shekarrizfard, M., Minet, L., Abrahamowicz, M., Baumgartner, J., Liu, R., Hatzopoulou, M., Weichenthal, S., 2020. Acute cardiovascular health effects in a panel study of personal exposure to traffic-related air pollutants and noise in Toronto, Canada. *Scientific Reports*. <https://doi.org/10.1038/s41598-020-73412-6>.
- Bruce-Low, S.S., Cotterrell, D., Jones, G.E., 2006. Heart rate variability during high ambient heat exposure. *Aviat Space Environ. Med.* 77 (9), 915–920.
- Bruce, U., Frances, S., Paul, C., Jeffrey, R.B., Karl, Z.L., Sanjay, R., Robert, D.B., 2005. Acute blood pressure responses in healthy adults during controlled air pollution exposures. *Environ. Health Perspect.* 113 (8), 1052–1055. <https://doi.org/10.1289/ehp.7785>.
- Cao, Q., Rui, G., Liang, Y., 2018. 5 pollution and the mortality due to lung cancer in China based on geographic weighted regression model. *BMC Publ. Health* 1–10. <https://doi.org/10.1186/s12889-018-5844-4>.
- Caplin, A., Ghandehari, M., Lim, C., Glimcher, P., Thurston, G., 2019. Advancing environmental exposure assessment science to benefit society. *Nature Communications*, 10(1) 1236. <https://doi.org/10.1038/s41467-019-09155-4>.
- Chen, S.-Y., Wu, C.-F., Lee, J.-H., Hoffmann, B., Peters, A., Brunekreef, B., Chu, D.-C., Chan, C.-C., 2015. Associations between long-term air pollutant exposures and blood pressure in elderly residents of Taipei city: a cross-sectional study. *Environ. Health Perspect.* 123 (8), 779–784.
- Dockery, D.W., Pope, C.A., Kanner, R.E., Villegas, G.M., Schwartz, J., 1999. Daily changes in oxygen saturation and pulse rate associated with particulate air pollution and barometric pressure. *Res. Rep. Health Eff. Inst.* 83(1) 1-28.
- Edgington, S., O'Sullivan, D.E., King, W.D., Lougheed, M.D., 2021. The effect of acute outdoor air pollution on peak expiratory flow in individuals with asthma: a systematic review and meta-analysis. *Environ. Res.* 192 (110296) <https://doi.org/10.1016/j.envres.2020.110296>.
- Ezzati, M., Dockery, D.W., 2009. Fine-particulate air pollution and life expectancy in the United States. *N. Engl. J. Med.* 360 (4), 376–386. <https://doi.org/10.1056/NEJMsa0805646>.
- Gold, D.R., Litonjua, A., Schwartz, J., Lovett, E., Larson, A., Nearing, B., Allen, G., Verrier, M., Cherry, R., Verrier, R., 2000. Ambient pollution and heart rate variability. *Circulation* 101 (11), 1267–1273. <https://doi.org/10.1161/01.CIR.101.11.1267>.
- Grundstein, A., Meentemeyer, V., Dowd, J., 2009. Maximum vehicle cabin temperatures under different meteorological conditions. *Int. J. Biometeorol.* 53 (3), 255–261. <https://doi.org/10.1007/s00484-009-0211-x>.
- Heal, M.R., Kumar, P., Harrison, R.M., 2012. Particles, air quality, policy and health. *Chem. Soc. Rev.* 41 (19), 6606–6630. <https://doi.org/10.1039/C2CS35076A>.
- Hong, C.L., F., K. S., David, S., Lawrence, B.W., Floyd, P., Mark, G., David, A., 2005. The association between fatal coronary heart disease and ambient particulate air pollution: are females at greater risk? *Environ. Health Perspect.* 113 (12), 1723–1729. <https://doi.org/10.1289/ehp.8190>.
- Ibald-Mulli, A., Timonen, K.L., Peters, A., Heinrich, J., Wölke, G., Lanki, T., Buzorius, G., Kreyling, W.G., de Hartog, J., Hoek, G., Brink, H.M., ten & Pekkanen, J., 2004. Effects of particulate air pollution on blood pressure and heart rate in subjects with cardiovascular disease: a multicenter approach. *Environ. Health Perspect.* 112 (3), 369–377. <http://www.jstor.org/stable/3435662>.
- Kateryna, F., Gudrun, W., Maria, F., Julia, D., Regina, H., Danny, H., Bente, O., Anna, O., Sviatlana, P., Johanna, P.N., Mette, S.J., Pekka, S., T., Kathrin, W., W., Inmaculada, X.W., Xavier, A., Rob, B., B., L., B. M., Barbara, H., 2014. Arterial blood pressure and long-term exposure to traffic-related air pollution: an analysis in the European study of cohorts for air pollution effects (ESCAPE). *Environ. Health Perspect.* 122 (9), 896–905. <https://doi.org/10.1289/ehp.130725>.
- Kupcickova, Z., Fecht, D., Ramakrishnan, R., Clark, C., Cai, Y.S., 2021. Road traffic noise and cardiovascular disease risk factors in UK Biobank. *Eur. Heart J.* 42 (21), 2072–2084. <https://doi.org/10.1093/eurheartj/ehab121>.
- Li, N., Chen, G., Liu, F., Mao, S., Liu, Y., Liu, S., Mao, Z., Lu, Y., Wang, C., Guo, Y., Xiang, H., Li, S., 2020. Associations between long-term exposure to air pollution and blood pressure and effect modifications by behavioral factors. *Environ. Res.* 182 (109109) <https://doi.org/10.1016/j.envres.2019.109109>.
- Lim, S., Barratt, B., Holliday, L., Griffiths, C.J., Mudway, I.S., 2021. Characterising professional drivers' exposure to traffic-related air pollution: evidence for reduction strategies from in-vehicle personal exposure monitoring. *Environ. Int.* 153 (106532) <https://doi.org/10.1016/j.envint.2021.106532>.
- Lipfert, F.W., Wyzga, R.E., 2008. On exposure and response relationships for health effects associated with exposure to vehicular traffic. *J. Expo. Sci. Environ. Epidemiol.* 18 (6), 588–599. <https://doi.org/10.1038/jes.2008.4>.
- Liu, L., Kauri, L.M., Mahmud, M., Weichenthal, S., Cakmak, S., Shutt, R., You, H., Thomson, E., Vincent, R., Kumarathasan, P., Broad, G., Dales, R., 2014. Exposure to air pollution near a steel plant and effects on cardiovascular physiology: a randomized crossover study. *Int. J. Hyg Environ. Health* 217 (2), 279–286. <https://doi.org/10.1016/j.ijheh.2013.06.007>.
- Longo, B.M., Rossignol, A., Green, J.B., 2008. Cardiorespiratory health effects associated with sulphurous volcanic air pollution. *Publ. Health* 122 (8), 809–820. <https://doi.org/10.1016/j.puhe.2007.09.017>.
- Madaniyazi, L., Zhou, Y., Li, S., Williams, G., Jaakkola, J.J.K., Liang, X., Liu, Y., Wu, S., Guo, Y., 2016. Outdoor temperature, heart rate and blood pressure in Chinese adults: effect modification by individual characteristics. *Scientific Reports*, 6(1) 21003. <https://doi.org/10.1038/srep21003>.
- McLaren, C., Null, J., Quinn, J., 2005. Heat stress from enclosed vehicles: moderate ambient temperatures cause significant temperature rise in enclosed vehicles. *Pediatrics* 116 (1), e109–e112. <https://doi.org/10.1542/peds.2004.2368>.
- Mette, S., Barbara, H., Martin, H., Matthias, K., Solvang, J.S., Jovanovic, A.Z., Anne, T., Kim, O., Ole, R.-N., 2012. Long-term exposure to traffic-related air pollution associated with blood pressure and self-reported hypertension in a Danish cohort. *Environ. Health Perspect.* 120 (3), 418–424. <https://doi.org/10.1289/ehp.1103631>.
- Miller, K.A., Siscovick, D.S., Sheppard, L., Shepherd, K., Sullivan, J.H., Anderson, G.L., Kaufman, J.D., 2007. Long-term exposure to air pollution and incidence of cardiovascular events in women. *N. Engl. J. Med.* 356 (5), 447–458. <https://doi.org/10.1056/NEJMoa054409>.
- Pangaribuan, M., Chuang, K.-J., Chuang, H.-C., 2019. Association between exposures to air pollution and biomarkers of cardiovascular disease in Northern Taiwan. *Atmos. Pollut. Res.* 10 (4), 1250–1259. <https://doi.org/10.1016/j.apr.2019.02.008>.
- Peters, A., Perz, S., Döring, A., Stieber, J., Koenig, W., Wichmann, H.E., 1999. Increases in heart rate during an air pollution episode. *Am. J. Epidemiol.* 150 (10), 1094–1098. <https://doi.org/10.1093/oxfordjournals.aje.a009934>.
- Peters, A., von Klot, S., Heier, M., Trentinaglia, I., 2004. Exposure to traffic and the onset of myocardial infarction. *N. Engl. J. Med.* 351 (17), 1721–1730. <https://doi.org/10.1056/NEJMoa040203>.
- Phung, D., Hien, T.T., Linh, H.N., Luong, L.M.T., Morawska, L., Chu, C., Binh, N.D., Thai, P.K., 2016. Air pollution and risk of respiratory and cardiovascular hospitalizations in the most populous city in Vietnam. *Sci. Total Environ.* 557, 322–330. <https://doi.org/10.1016/j.scitotenv.2016.03.070>, 558.
- Pope, C.A.I., Dockery, D.W., Kanner, R.E., Villegas, G.M., Schwartz, J., 1999. Oxygen saturation, pulse rate, and particulate air pollution. *Am. J. Respir. Crit. Care Med.* 159 (2), 365–372. <https://doi.org/10.1164/ajrccm.159.2.9702103>.
- Reis, L.A., Drouet, L., Tavoni, M., 2022. Internalising health-economic impacts of air pollution into climate policy: a global modelling study. *Lancet Planet. Health* 6 (1), e40–e48. [https://doi.org/10.1016/S2542-5196\(21\)00259-X](https://doi.org/10.1016/S2542-5196(21)00259-X).
- Ren, C., O'Neill, M.S., Park, S.K., Sparrow, D., Vokonas, P., Schwartz, J., 2011. Ambient temperature, air pollution, and heart rate variability in an aging population. *Am. J. Epidemiol.* 173 (9), 1013–1021. <https://doi.org/10.1093/aje/kwq477>.
- Roh, M., Jeon, S., Kim, S., Yu, S., Heshmati, A., Kim, S., 2020. Modeling air pollutant emissions in the provincial level road transportation sector in Korea: a case study of the zero-emission vehicle subsidy. *Energies* 5 (3999). <https://doi.org/10.3390/en13153999>.
- Schikowski, T., Sugiri, D., Ranft, U., Gehring, U., Heinrich, J., Wichmann, H.-E., Krämer, U., 2005. Long-term air pollution exposure and living close to busy roads are associated with COPD in women. *Respiratory Research*. <https://doi.org/10.1186/1465-9921-6-152>.
- Setton, E., Marshall, J.D., Brauer, M., Lundquist, K.R., Hystad, P., Keller, P., Cloutier-Fisher, D., 2011. The impact of daily mobility on exposure to traffic-related air

- pollution and health effect estimates. *J. Expo. Sci. Environ. Epidemiol.* 21 (1), 42–48. <https://doi.org/10.1038/jes.2010.14>.
- Stafoggia, M., Oftedal, B., Chen, J., Rodopoulou, S., Renzi, M., Atkinson, R.W., Bauwelinck, M., Klompaker, J.O., Mehta, A., Vienneau, D., Andersen, Z.J., Bellander, T., Brandt, J., Cesaroni, G., de Hoogh, K., Fecht, D., Gulliver, J., Hertel, O., Hoffmann, B., Janssen, N.A.H., 2022. Long-term exposure to low ambient air pollution concentrations and mortality among 28 million people: results from seven large European cohorts within the ELAPSE project. *Lancet Planet. Health* 6 (1), e9–e18. [https://doi.org/10.1016/S2542-5196\(21\)00277-1](https://doi.org/10.1016/S2542-5196(21)00277-1).
- Suarez-Bertoa, R., Valverde, V., Pavlovic, J., Clairotte, M., Selleri, T., Franco, V., Kregar, Z., Astorga, C., 2021. On-road emissions of Euro 6d-TEMP passenger cars on Alpine routes during the winter period. *Environ. Sci. J. Integr. Environ. Res. Atmosphere* 1 (3), 125–139. <https://doi.org/10.1039/D0EA00010H>.
- Tartakovsky, L., Baibikov, V., Czerwinski, J., Gutman, M., Kasper, M., Popescu, D., Veinblat, M., Zvirin, Y., 2013. In-vehicle particle air pollution and its mitigation. *Atmos. Environ.* 64, 320–328. <https://doi.org/10.1016/j.atmosenv.2012.10.003>.
- Walsh, M.P., 2008. Ancillary benefits for climate change mitigation and air pollution control in the world's motor vehicle fleets. *Annu. Rev. Publ. Health* 29 (1), 1–9. <https://doi.org/10.1146/annurev.publhealth.29.091307.083257>.
- WHO, 2010. A healthy lifestyle - WHO recommendations. <https://www.who.int/europe/news-room/fact-sheets/item/a-healthy-lifestyle-who-recommendations>.
- Wu, T., Cui, Y., Lian, A., Tian, Y., Li, R., Liu, X., Yan, J., Xue, Y., Liu, H., Wu, B., 2022. Vehicle emissions of primary air pollutants from 2009 to 2019 and projection for the 14th Five-Year Plan period in Beijing, China. *J. Environ. Sci.* 124, 513–521. <https://doi.org/10.1016/j.jes.2021.11.038>.
- Zhang, A.L., Balmes, J.R., Lutzker, L., Mann, J.K., Margolis, H.G., Tyner, T., Holland, N., Noth, E.M., Lurmann, F., Hammond, S.K., Holm, S.M., 2022. Traffic-related air pollution, biomarkers of metabolic dysfunction, oxidative stress, and CCl6 in children. *J. Expo. Sci. Environ. Epidemiol.* 32 (4), 530–537. <https://doi.org/10.1038/s41370-021-00378-6>.

Effect of short-term exposures to traffic air pollution and temperature on heart rate among healthy female students during commuting

ORIGINALITY REPORT

11%

SIMILARITY INDEX

9%

INTERNET SOURCES

9%

PUBLICATIONS

%

STUDENT PAPERS

PRIMARY SOURCES

1	www.energy.ca.gov Internet Source	2%
2	Xiangmei (May) Wu, Michael G. Apte, Deborah H. Bennett. "Indoor Particle Levels in Small- and Medium-Sized Commercial Buildings in California", <i>Environmental Science & Technology</i> , 2012 Publication	2%
3	docplayer.net Internet Source	1%
4	indico.cells.es Internet Source	1%
5	Mette Sørensen, Barbara Hoffmann, Martin Hvidberg, Matthias Ketzel et al. "Long-Term Exposure to Traffic-Related Air Pollution Associated with Blood Pressure and Self-Reported Hypertension in a Danish Cohort", <i>Environmental Health Perspectives</i> , 2012 Publication	1%

6	repositorio.unican.es Internet Source	1 %
7	www.ncbi.nlm.nih.gov Internet Source	1 %
8	eprints.lums.ac.ir Internet Source	1 %
9	www.acarindex.com Internet Source	1 %
10	vufind.katalog.k.utb.cz Internet Source	1 %
11	www.helmholtz-munich.de Internet Source	1 %

Exclude quotes On

Exclude bibliography On

Exclude matches < 1%